

# Additional Commentary on the Detection and Quantification of Plastic Micro- and Nanoparticles in Tea Samples

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**Abstract:** The study of plastic particles, particularly those in the micro-, sub-micro-, and nano-size ranges, within food and beverages has gained increasing interest within recent years. However, many analytical techniques have limits of detection which hinder their use for the study of these particles in these sample matrices. In addition, remaining contaminants from the matrices can interfere with the signals from plastic particles. Thus, great care must be given to sample preparation and data interpretation to ensure accurate results. This study proposes the use of sample purification through chemical digestion protocols to facilitate the study of plastic particles present in tea samples, and serves to highlight technical limitations which must be overcome in future studies.

**Keywords:** Beverages · Food · Microplastic · Nanoplastic



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## 1. Introduction

Of the 368 million tons of plastics produced in 2020, 39.6% (approximately 145.7 million tons) were utilized in packaging materials.<sup>[1]</sup> Commonly, such packaging is utilized to preserve food and beverage products meant for human consumption. Many recent studies report the subsequent transmission of plastic particles with sizes in the micro- (5 mm – 1 µm), sub-micro (1 µm – 100 nm), and even nano- (<100 nm) ranges into the products that are contained within such plastic packaging. However, to date, the study of nanoplastic particles in analytically complex sample matrices (*i.e.* those that contain high amounts of additional organic or inorganic matter) is plagued by a multitude of challenges which arise from their predominantly carbon-based chemical structure, low densities, small sizes, and low estimated concentrations within relevant samples. Thus, studies tend to focus on larger microplastic particle presence, and have reported their detection in a variety of food and beverage products such as bottled water<sup>[2]</sup> and beers.<sup>[3]</sup> While this data serves to highlight the importance of quantifying the presence of plastic particles within products meant for human consumption, it fails to ac-

curately represent the most analytically challenging fraction of plastic particles: nanoplastics.

Research groups such as Hernandez *et al.*<sup>[4]</sup> have attempted to address this knowledge gap through the study of micro-, sub-micro-, and nanoplastic particles reported to leach from teabags into tea after the hydrolytic degradation<sup>[5]</sup> of the polymer chains present in the plastics that the teabags are composed of. However, many of the standard techniques utilized in the literature have limitations that hinder the accurate collection and interpretation of data; such as the limit of detection commonly reported for Fourier transform infrared spectroscopy (FTIR) being orders of magnitude larger (*i.e.* ~10 µm minimum particle size<sup>[6]</sup>) than the nanoplastic particles of interest or the difficulty differentiating true plastic particles from other organic materials in scanning electron microscopy (SEM) images.<sup>[7]</sup> Such concerns have previously been expressed by Busse *et al.*, who highlighted the potential for the results reported by Hernandez *et al.* to be skewed by the presence of contaminants such as oligomers that would not be differentiated from true plastic particles by FTIR due to the area averaging quality of the measurements, or in SEM images due to oligomer crystallization upon drying.<sup>[7]</sup> Thus, the aim of this study is to further probe the limit of these techniques for the detection of plastic particles in analytically complex sample matrices such as foods and beverages, and to attempt to augment the data one can obtain through the introduction of additional controls and purification steps like the chemical digestion of contaminants previously reported in the literature.<sup>[8]</sup>

## 2. Results and Interpretation

Commercial teabag samples were prepared through cutting and rinsing and the recovered tea leaves were steeped as a control as reported by Hernandez *et al.*<sup>[4]</sup> Additional controls were prepared utilizing plastic teabags, which had never contained tea leaves, and are henceforth referred to as ‘empty control’. SEM images obtained using a Mira3 LM FE scanning electron microscope

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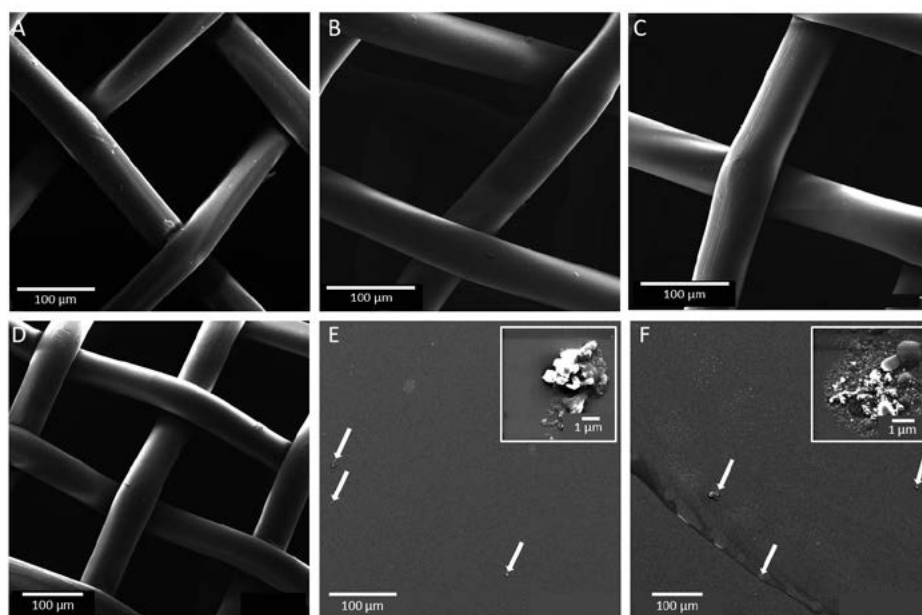
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(Tescan, Czech Republic) for the teabags before and after the steeping protocol (Fig. 1) yielded similar results to those reported by Hernandez *et al.*<sup>[4]</sup> Namely, small particles could be found on the surface of non-steeped teabags which were not present on the surface of steeped teabags, and surface roughness could be seen for steeped teabags. However, it was noted that areas with higher surface roughness could also be observed in the non-steeped teabags and that particles could still be observed on the surface of steeped teabags. Thus, the potential that some defects may be present from manufacturing should be considered. Additionally, SEM images of procedural blank samples (*i.e.* MilliQ water which was boiled, filtered, dialyzed, digested, and handled in the same manner as true samples) which were prepared under the same conditions reported by Hernandez *et al.* were shown to contain trace amounts of particle contaminants (*i.e.* from atmospheric deposition) when imaged at low magnification (Fig. 1). Thus, while the cut teabag leachates showed what appeared to be particles in size ranges similar to those reported by Hernandez *et al.*,<sup>[4]</sup> the exact composition of these particle-like materials requires further investigation. The need for such investigation is further highlighted by Busse *et al.* in their recent commentary,<sup>[7]</sup> where they discuss the potential for the particle-like materials present in SEM images to be crystallized oligomers, or contaminants as a result of sample handling and preparation.

Fig. 1. SEM images of the commercial teabags before (A) and after (B) steeping, the empty nylon control teabags before (C) and after (D) steeping, the initial procedural blank leachate (*i.e.* after boiling, filtration, and dialysis) (E) and the procedural blank after a 7-day KOH digestion (F) show the presence of trace amounts of particles (indicated with white arrows). Insets for the procedural blanks show particles at a higher magnification.



With this in mind, ATR-FTIR spectra were collected for the commercial teabags in addition to multiple controls (Fig. 2; Table 1) using a Perkin Elmer universal attenuated total reflection (UATR) accessory and pressure arm. The commercial teabag spectra and the empty teabag control matched with a Nylon 6 reference polymer spectrum present in the Perkin Elmer Spectra 10 reference database (Table 2). However, the spectra for the cut commercial teabag leachates ('Commercial Leachate' in Fig. 2) had extra peaks that were not observed in the spectra obtained from the non-steeped commercial teabags. These peaks are present even after the completion of the full purification process (*i.e.* filtration with a cellulose 2.5  $\mu\text{m}$  cutoff Whatmann filter, 5–7 day dialysis with a 14 kDa cutoff membrane from Carl Roth GmbH + Co). Some of these additional peaks (*e.g.* at  $\sim 1,012\text{ cm}^{-1}$ ) have been reported by Oh *et al.*<sup>[10]</sup> to appear in the FTIR spectra obtained for cellulose as a result of the presence of  $-\text{OH}$  bonds in the cellulose backbone and may be attributed to the presence of contaminants from the filter paper, atmospheric deposition, or organic matter left in the teabags after the rinsing protocol. Other additional peaks observed are reported in a study by Cai *et al.* to be present in tea samples; with specific bands near  $1,600\text{ cm}^{-1}$  reported to be the result of ring stretching from sugars and/or  $\text{N}-\text{H}$  and  $\text{C}-\text{O}$  bond vibrations from carboxyl groups, near  $1,000\text{--}1,100\text{ cm}^{-1}$  reported to be characteristic of the O-substituted glucose residues in

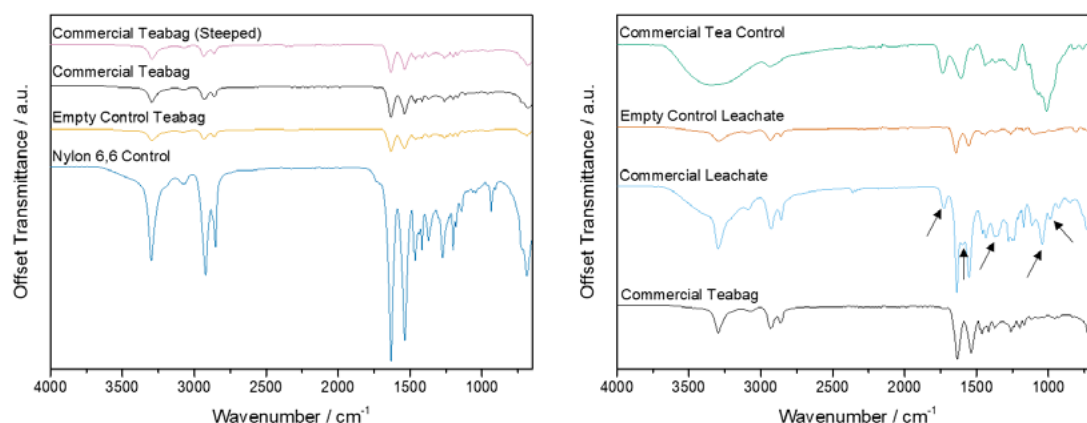


Fig. 2. FTIR data for the representative commercially available teabag and its leachate compared to spectra from relevant controls. All spectra were obtained by averaging 10 scans over the wavenumber range of  $650\text{--}4,000\text{ cm}^{-1}$  with a spectral resolution of  $4\text{ cm}^{-1}$ . Extra peaks present in the commercial teabag leachate are indicated with black arrows.

Table 1. Peak assignments for peaks anticipated to be in the FTIR spectra obtained from samples that are Nylon- based.<sup>[9]</sup>

Wavenumber (cm <sup>-1</sup> )	Samples Containing Peak	Assignment
3,296	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	N–H stretching vibration
3,000–2,780	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	C–H symmetric stretching vibration
1,640	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	C=O stretching vibration (amide I)
1,540	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	N–H bending; C–N stretching vibrations (amide II)
1,480–1,400	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	CH <sub>2</sub> symmetric (scissor) deformation
1,350–1,200	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	N–H deformation (amide III)
1,170	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	CONH skeletal vibrations
1,124–1,020	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	C–C and CONH skeletal vibrations
960–930	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	CONH skeletal vibrations
~690	Nylon 6,6 Control; Empty Teabag Control and Leachate; Commercial teabag	N–H wagging

Table 2: A summary of the similarity of chemical fingerprints of various tested materials obtained through automated correlation analysis.

Sample 1	Sample 2	Correlation
Empty Control Teabag	Pure Nylon Reference Spectra	96%
Empty Control Teabag	Empty Control Leachate	79%
Empty Control Leachate	Pure Nylon Reference Spectra	75%
Commercial Teabag	Pure Nylon Reference Spectra	93%
Commercial Teabag	Commercial Teabag Leachate	30%
Commercial Teabag Leachate	Commercial Steeped Tea Control	44%
Empty Control Teabag	Commercial Teabag	96%
Commercial Teabag Leachate	Empty Control Leachate	18%
Empty Control Leachate 1:1 KOH 7 Days	Empty Control Teabag	14%

tea polysaccharides, and within the ranges of 1,700 to 1,460 cm<sup>-1</sup> and 800 to 500 cm<sup>-1</sup> as the result of various tea polysaccharides.<sup>[11]</sup> Such results appear to indicate the presence of small tea leaf frag-

ments or other contaminants within the samples even after the washing, filtration, and dialysis steps utilized in the sample purification protocol.

In order to be able to more accurately determine the number of particles present in the cut teabag leachates that were truly plastic-based, steps were made to develop a protocol for the chemical digestion of particulate organic matter from tea leaves. As it was previously reported in the literature that potassium hydroxide (KOH) gave high digestion efficiencies for animal tissue with minimal damage to microplastic particles<sup>[8a]</sup> and that hydrogen peroxide (H<sub>2</sub>O<sub>2</sub>) yielded the highest digestion efficiencies for vegetal matter<sup>[8b]</sup> these were the chemicals selected for initial testing. For both KOH and H<sub>2</sub>O<sub>2</sub>, a digestion period of 48 hours was shown to be insufficient for the removal of organic matter present in tea samples. The sample digested with H<sub>2</sub>O<sub>2</sub> for 7 days also yielded unsatisfactory results, but after a 7-day digestion with KOH it appeared that some organic matter may have been digested out of the sample. However, key peaks found in the initial tea control spectra could still be seen, albeit at lower intensities, in the chemical fingerprint of the digested samples. Additionally, when the leachate obtained from steeping the empty control teabags was digested under these conditions, a severe drop in the correlation value with the teabag was observed (Table 2; Fig. 3). This indicates that attempts to chemically degrade the components from tea will additionally degrade the plastic-based oligomers, but may also damage any plastic particles present within the sample. Such findings have been reported frequently in the literature for plastic particles composed of Nylon (*i.e.* polyamide) and poly(ethylene terephthalate)<sup>[8a,12]</sup> due to the potential for the polymer chains that compose these materials to hydrolyze in aqueous environments.<sup>[5]</sup>

### 3. Conclusions

For the teabag types analyzed during the current study, it is likely that the leachate obtained from steeping the cut, rinsed commercial teabags is a mixture of plastic-based materials and other

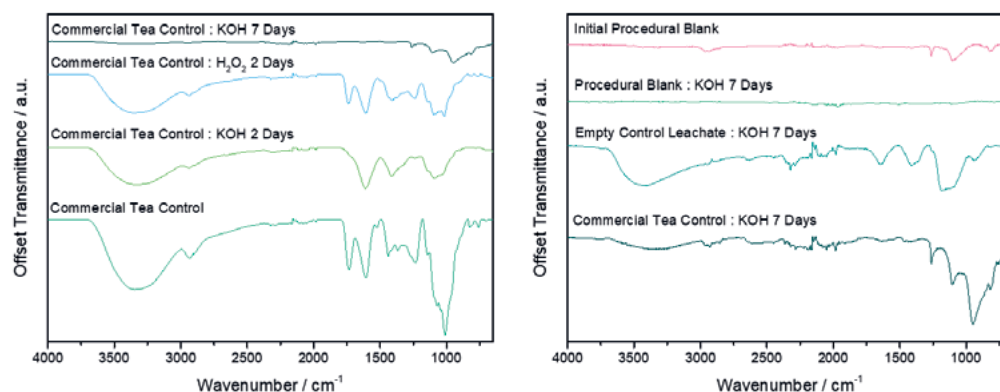


Fig. 3. FTIR data for the representative commercially available tea control and the empty control teabag leachate after chemical digestion. Additional comparisons are made to procedural blank data collected before and after the most promising chemical digestion protocol. All spectra were obtained by averaging 10 scans over the wavenumber range of 650–4,000  $\text{cm}^{-1}$  with a spectral resolution of 4  $\text{cm}^{-1}$ .

organic contaminants such as tea leaf-based particulate matter or dust. This hypothesis is supported by the literature<sup>[7]</sup> in addition to poor matches between the FTIR spectra of the teabags and the leachates, and the presence of trace amounts of particulate matter detected within the SEM imaging controls. However, the techniques tested in this study were shown to be incapable of resolving the chemical contents of the leachate samples on a single particle or single material basis, and the chemical digestions were shown to degrade organic contaminants in addition to potential plastic-based contaminants. Thus, while a qualitative sample analysis could be achieved, no quantitative data can be accurately reported. The authors encourage future studies to be conducted only after careful consideration of the data presented within the field, and with the plans for further optimization of sample pre-processing (e.g. digestion with enzymes to ensure no plastic degradation<sup>[13]</sup>) and/or analytics (e.g. through the introduction of more analytically sensitive techniques such as SERS<sup>[14]</sup>), which may allow for the analysis of samples in a single-particle or single material basis.

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- [1] PlasticsEurope, 'Plastics – the Facts 2020 an Analysis of European Plastics Production, Demand and Waste Data' Brussels – Belgium, **2020**.
- [2] S. A. Mason, V. G. Welch, J. Neratko, *Front. Chem.* **2018**, *6*, <https://doi.org/10.3389/fchem.2018.00407>.
- [3] G. Liebezeit, E. Liebezeit, *Food Add. Contamin. A* **2014**, *31*, 1574, <https://doi.org/10.1080/19440049.2014.945099>.
- [4] L. M. Hernandez, E. G. Xu, H. C. E. Larsson, R. Tahara, V. B. Maisuria, N. Tufenkji, *Environ. Sci. Technol.* **2019**, *53*, 12300, <https://doi.org/10.1021/acs.est.9b02540>.
- [5] B. Gewert, M. M. Plassmann, M. MacLeod, *Environ. Sci. Proc. Impacts* **2015**, *17*, 1513, <https://doi.org/10.1039/C5EM00207A>.
- [6] A. K ppler, D. Fischer, S. Oberbeckmann, G. Schernewski, M. Labrenz, K.-J. Eichhorn, B. Voit, *Anal. Bioanal. Chem.* **2016**, *408*, 8377, <https://doi.org/10.1007/s00216-016-9956-3>.
- [7] K. Busse, I. Ebner, H.-U. Humpf, N. Ivleva, A. K ppler, B. E. O bmann, D. Schymanski, *Environ. Sci. Technol.* **2020**, *54*, 14134, <https://doi.org/10.1021/acs.est.0c03182>.

- [8] a) A. Karami, A. Golieskardi, C. K. Choo, N. Romano, Y. B. Ho, B. Salamatnia, *Sci. Total Environ.* **2017**, *578*, 485, <https://doi.org/10.1016/j.scitotenv.2016.10.213>; b) A. Herrera, P. Garrido-Amador, I. Mart nez, M. D. Samper, J. L pez-Mart nez, M. G mez, T. T. Packard, *Marine Poll. Bull.* **2018**, *129*, 61, <https://doi.org/10.1016/j.marpolbul.2018.02.015>.
- [9] a) G. Socrates, 'Infrared and Raman characteristic group frequencies: tables and charts', John Wiley & Sons, **2004**; b) I. Noda, A. E. Dowrey, J. L. Haynes, C. Marcott, in 'Physical Properties of Polymers Handbook', Ed. J. E. Mark, Springer New York, New York, NY, **2007**, p. 395, [https://doi.org/10.1007/978-0-387-69002-5\\_22](https://doi.org/10.1007/978-0-387-69002-5_22); c) G. Rotter, H. Ishida, *J. Polym. Sci. B* **1992**, *30*, 489, <https://doi.org/10.1002/polb.1992.090300508>.
- [10] S. Y. Oh, D. I. Yoo, Y. Shin, G. Seo, *Carbohydr. Res.* **2005**, *340*, 417, <https://doi.org/10.1016/j.carres.2004.11.027>.
- [11] J.-X. Cai, Y.-F. Wang, X.-G. Xi, H. Li, X.-L. Wei, *Int. J. Biol. Macromol.* **2015**, *78*, 439, <https://doi.org/10.1016/j.ijbiomac.2015.03.025>.
- [12] a) F. Ribeiro, E. D. Okoffo, J. W. O'Brien, S. Fraissinet-Tachet, S. O'Brien, M. Gallen, S. Samanipour, S. Kaserzon, J. F. Mueller, T. Galloway, K. V. Thomas, *Environ. Sci. Technol.* **2020**, *54*, 9408, <https://doi.org/10.1021/acs.est.0c02337>; b) M. Scheurer, M. Bigalke, *Environ. Sci. Technol.* **2018**, *52*, 3591, <https://doi.org/10.1021/acs.est.7b06003>.
- [13] a) A. Dehaut, A.-L. Cassone, L. Fr re, L. Hermabessiere, C. Himber, E. Rinnert, G. Riviere, C. Lambert, P. Soudant, A. Huvet, G. Duflos, I. Paul-Pont, *Environ. Poll.* **2016**, *215*, 223, <https://doi.org/10.1016/j.envpol.2016.05.018>; b) W. Courtene-Jones, B. Quinn, F. Murphy, S. F. Gary, B. E. Narayanaswamy, *Anal. Meth.* **2017**, *9*, 1437, <https://doi.org/10.1039/C6AY02343F>.
- [14] J. Langer, D. Jimenez de Aberasturi, J. Aizpurua, R. A. Alvarez-Puebla, B. Augu e, J. J. Baumberg, G. C. Bazan, S. E. J. Bell, A. Boisen, A. G. Brolo, J. Choo, D. Cialla-May, V. Deckert, L. Fabris, K. Faulds, F. J. Garc a de Abajo, R. Goodacre, D. Graham, A. J. Haes, C. L. Haynes, C. Huck, T. Itoh, M. K ll, J. Kneipp, N. A. Kotov, H. Kuang, E. C. Le Ru, H. K. Lee, J.-F. Li, X. Y. Ling, S. A. Maier, T. Mayerh fer, M. Moskovits, K. Murakoshi, J.-M. Nam, S. Nie, Y. Ozaki, I. Pastoriza-Santos, J. Perez-Juste, J. Popp, A. Pucci, S. Reich, B. Ren, G. C. Schatz, T. Shegai, S. Schl cker, L.-L. Tay, K. G. Thomas, Z.-Q. Tian, R. P. Van Duyne, T. Vo-Dinh, Y. Wang, K. A. Willets, C. Xu, H. Xu, Y. Xu, Y. S. Yamamoto, B. Zhao, L. M. Liz-Marz n, *ACS Nano* **2020**, *14*, 28, <https://doi.org/10.1021/acs.nano.9b04224>.

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